

Optimization of a Novel Procedure for Determination of VOCs in Water and Human Urine Samples Based on SBSE Coupled with TD-GC–HRMS

Natalia Jakubowska^{1,*}, Bernhard Henkelmann², Karl-Werner Schramm², and Jacek Namiesnik³

¹Department of Analytical Chemistry, Chemical Faculty, Gdańsk University of Technology (GUT), 11/12 G. Narutowicza Street, 80-952 Gdansk, Poland; ²Helmholtz Zentrum München, Institute of Ecological Chemistry, Ingolstaedter Landstr. 1, D-85764 Neuherberg, Germany; and ³Gdansk University of Technology, 80-233 Gdansk

Abstract

In this study, stir-bar sorptive extraction and thermal desorption followed by gas chromatography coupled with high resolution mass spectrometry was applied for determination of halo-organic compounds (bromodichloromethane, dibromochloromethane, bromoform, and tetrachloroethylene) in water and human urine samples. Time of extraction and stirring speed were optimized. The results show that the optimum extraction time is 30 min with 600 rpm of stirring speed with Twister of 20 mm in length and 1.0-mm film thickness of PDMS (126 μ L). The calibration curves, limits of detection and quantification for all compounds were calculated. This procedure is characterized by very low limits of detection and quantitation: lower than 0.0017 μ g/L and good repeatability for all four volatile compounds. This new analytical procedure was identified to be easy, reliable, sensitive, and requires only small amounts of sample. It can constitute a good alternative to well-known procedures based on application of head space and gas chromatography coupled with electron capture detection.

Introduction

Volatile halogenated compounds [VOXs, also known as halogenated volatile organic compounds (HVOCs)] such as bromodichloromethane (CHCl_2Br), dibromochloromethane (CHClBr_2), tribromomethane (bromoform, CHBr_3), and tetrachloroethene (tetrachloroethylene, perchloroethylene, C_2Cl_4) belong to the most important pollutants of indoor and workplace air. People are exposed to these compounds in their homes and workplaces during varying activities of everyday life. These compounds are present in drinking water and water of swimming pools due to chlorination for disinfection purposes. They are also widely employed in industry as degreasing agents (1) and even more commonly are used for the dry-cleaning of clothes (2–4).

C_2Cl_4 has been classified as group 2A carcinogen by the International Cancer Research Institution and CHCl_2Br as group 2B carcinogen, which means that they are possibly carcinogenic to humans (5).

Halogenated compounds can enter the human body by many different routes; for example, by inhalation, dermal contact, or inadvertent ingestion from hand-to-mouth contact. After intake, the chemicals may enter the bloodstream, and in the body they can be accumulated or are excreted, usually via urine (in non-metabolised form) (6,7).

Liquid samples usually require special treatment prior to the final analysis [e.g., by gas chromatography (GC)]. The sample preparation techniques commonly used in water analysis for the content of HVOCs are gas, sorbent, solvent, and membrane extraction. To prepare liquid biological samples for GC analysis, one must take into consideration the nature of the matrix, the method of sample introduction into a GC column, and the limited quantity of the sample. Due to these facts, urine samples have become of great interest for analysts. Special attention has recently been paid to the use of so-called solvent-free analyte isolation and/or enrichment techniques, which can be attributed to the widespread use of green analytical chemistry (8). In practice, various implementations of the headspace (HS) technique are most often used for this purpose, with static HS being the most popular. However, HS technique is non-selective towards volatile compounds and can require long sampling times. Still, analytical chemists feel the need to search for a new methodological and instrumental approach. Stir-bar sorptive extraction (SBSE) can constitute the technique of choice for this task (9,10). SBSE is a novel technique based on the same principles as the well-known SPME technique. Partitioning coefficient of the solutes between the silicone phase and the aqueous phase has been evaluated for the enrichment of volatile organic compounds from water and biological fluid samples. In the case of SBSE technique, stir-bars were coated with a 50–250 times larger amount of polydimethylsiloxane (PDMS) layer than in SPME technique,

*Author to whom correspondence should be addressed: email nataliajakubowska6@wp.pl.

which increases the preconcentration capacity and recovery, and decreases the limit of detection (LOD). Another advantage offered by this technique is the fact that the stir-bar does not have to be extensively dried before the desorption process, thus less of the volatile compounds are lost during this step (11–21).

The aim of this study was to evaluate SBSE technique, followed by thermal desorption and GC-HRMS analysis, for the determination of CHCl_2Br , CHClBr_2 , CHBr_3 , and C_2Cl_4 in water and human urine samples. Stirring rate and time of extraction were optimized. Basic metrological parameters such as linearity, LOD, and limit of quantification (LOQ) were calculated. Isotope dilution was used for the analysis of all samples, quality controls (QC), and standards. The use of the isotope dilution technique increased the precision and accuracy of the analysis.

Experimental

Reagents and analytical standards

Standards for optimization and calibration of the SBSE-TD-GC-HRMS procedure are as follows: CDBr_3 was purchased from Sigma-Aldrich (Steinheim, Germany); CHCl_2Br , CHClBr_2 , CHBr_3 , and C_2Cl_4 (200.00 mg/dm³, 5000.00 mg/dm³) were purchased from Supelco (Bellefonte, PA).

CH_3OH for chromatography was obtained from Merck (Darmstadt, Germany). "Zero water" (level of total organic carbon 1–4 $\mu\text{g/L}$ C) was produced by a Milli-Q Millipore system (Molsheim, France).

Theoretical recovery of VOX

Table I shows the $\log K_{o/w}$ and theoretical recovery values of four halo-organic compounds investigated in this work. The theoretical recovery (TR) was calculated by applying the following formula:

$$\text{TR} = (K_{o/w}/\beta)/(1 + K_{o/w}/\beta) = 1/(\beta/K_{o/w} + 1) \quad \text{Eq. 1}$$

where $\beta = V_s/V_{\text{PDMS}}$, V_{PDMS} is the volume of PDMS ($V_{\text{PDMS}} = 126 \mu\text{L}$), and V_s is the volume of the sample (water or urine).

Water and human urine samples

In the sampling of water, the basic rule is to fill up a container fully (no HS) and to keep it at about 4°C, protected from possible contamination. The isotope dilution method was used for the water samples. Various quantities of the analytes (0.5, 1, 2, 5, and 10 $\mu\text{g/L}$) and an equal amount of deuterated bromoform (2 $\mu\text{g/L}$) were added to each sample of millipore water (10 mL), followed by the analysis using the SBSE-TD-GC-HRMS procedure.

Urine samples should be collected without HS stored at ~4°C and analyzed within 24 h. One urine sample was collected from a volunteer and divided into five aliquots of equal volume. One of them was used as a blank, spiked with the CDBr_3 standard only, and analyzed directly for the content of organohalogen compounds. The other four aliquots were spiked with the same quantities of the analytes (2 $\mu\text{g/L}$) and the same amount of deuterated bromoform (2 $\mu\text{g/L}$), respectively. Every urine sample was analyzed using the SBSE-TD-GC-HRMS procedure.

Conditioning of the coated stir bars

The coated stir-bars (Twisters) for sorptive extraction were obtained from Gerstel (Gerstel GmbH, Mulheim an der Ruhr, Germany). Twisters of 20 mm of length and coated with a 1.0-mm thick film of PDMS (126 μL) were conditioned prior to the first and after each analysis as follows: used Twisters were placed into a clean 100-mL flask containing a 1:1 mixture of dichloromethane and methanol and shaken for 30 min on a rotating shaking machine to clean the PDMS phase. After 30 min, the solvent mixture was changed for a fresh one and shaken for another 30 min. The Twisters were removed from the solvent and dried for a short time on a clean surface at room temperature; afterwards, they were placed into clean TDS tubes and conditioned in a Gerstel tube conditioner at 300°C at a flow rate of helium of 100 mL/min for 1 h. After cooling down, the Twisters were placed into clean screw cap vials. After the stir bars were conditioned, no memory effect was observed.

Instrumentation

All analyses were performed by thermal desorption GC-MS. An Agilent GC 5890 Series II (Santa Clara, CA) was equipped with an autosampler Gerstel MPS 2 (Mulheim, Germany), a tray for 98 desorption tubes Gerstel VT98t, and a desorption unit Gerstel TDU which was coupled to a cold injection system Gerstel CIS 3. The Gerstel MAsTer software was used to control and set the parameters for the autosampler, desorption unit, and the injection system. The desorption (TDU to CIS) and the injection (CIS to column) were both performed in splitless mode at a helium flow of 70 mL/min. A liner filled with glass wool was installed in the CIS.

The temperature for desorption of the analytes was programmed from 30°C to 200°C at 60°C/min and a final hold of 5 min. The desorbed analytes were trapped in the CIS 3 at -100°C and afterwards heated to 200°C at 12°C/s for injection onto the GC column.

The GC system was connected to a high resolution mass spec-

Table I. Values of $\log K_{o/w}$ and Theoretical Recoveries of CHCl_2Br , CHClBr_2 , CHBr_3 , and C_2Cl_4

Compound	$\log K_{o/w}$	Sample volume (mL)	Theoretical recovery (%)
CHCl_2Br	2.00	2	86.3
		5	71.6
		10	55.8
		20	38.7
C_2Cl_4	3.40	2	99.37
		5	98.45
		10	96.94
		20	94.06
CHClBr_2	2.16	2	90.1
		5	78.5
		10	64.6
		20	47.7
CHBr_3	2.38	2	93.8
		5	85.8
		10	75.1
		20	60.2

trometer Thermo Scientific MAT 95 (Bremen, Germany) and a Restek Rtx-CL Pesticides 2 capillary column (Bellefonte, PA) (30 m × 0.25 mm i.d., 0.20- μ m film thickness) was used for the chromatographic separation. The GC temperature program was as follows: 30°C, 5 min; 12°C/min to 95°C, 2 min; 25°C/min to 200°C, 5 min. Helium served as carrier gas at a head pressure of 16 psi.

The MS was operated in SIM mode at a resolution of > 7000. The temperature of the ion source was 260°C. The two most intense ions of the molecular ion cluster or a high abundant fragment ion cluster for each compound were monitored, as summarized in Table II.

SBSE-TD-HRGC-HRMS procedure for determination of VOX in liquid samples

10 mL of sample (water or urine) and the standard solutions were pipetted into a special 10-mL THM flask from Supelco (Bellefonte, PA). A Twister was placed into the sample and stirred for 30 min at 600 rpm. After sampling, the stir-bar was taken out of the vial with tweezers and shortly dipped on a clean paper tissue to remove residual water droplets. The Twister was finally placed into an empty desorption liner of 60 mm length and 6 mm i.d. The desorbed analytes were detected as described earlier. In Figure 1, the general scheme of the whole procedure is depicted.

Results

Instrumental operating conditions

In a first approach, the GC–HRMS conditions including oven temperature, thermal desorption program, and retention time characteristic were evaluated. Instrumental optimization was

performed with CHCl_2Br , CHClBr_2 , CHBr_3 , and C_2Cl_4 standard solutions, which were directly spiked onto the stir bar by use of a zero dead-volume syringe (1 μL of a 200 mg/L solution).

Optimization of the SBSE

Sample volume, time of SBSE extraction, and stirring speed were evaluated to achieve the best overall analytical conditions.

All extraction experiments were carried out in special 10-mL screw cap vials, which can be filled up to a minimum of HS and provide a diameter wide enough to stir the coated bars in it. With a sample volume of 10 mL, a total recovery of the analytes of at minimum 56% is expected (Table I), resulting in a good performance of the method.

Optimization studies were carried out in water samples spiked at the 10 $\mu\text{g/L}$ level. The important parameters extraction time (30, 60, and 90 min) and stirring speed (500, 600, 800, and 1000 rpm) were optimized. Additionally, stir-bars having a PDMS phase volume of 126 μL were chosen because a higher extraction capacity is attained.

The ideal enrichment time is the one in which the amount of compounds detected reaches a maximum, and any subsequent increment in time does not result in a higher compound signal. When an application involves a mixture of compounds with multiple functionalities, different maxima are expected and a compromise is needed for choosing the enrichment time. For these experiments, 30, 45, and 60 min and 500, 600, 800, and 1000 rpm were evaluated. The results demonstrated that at 30 min a good detection capability was obtained for all four compounds and the deuterated CHBr_3 . It was observed that the peak area for each compound was relatively the same with increasing time, and at 600 rpm the best response for the standard was observed. (results for CHCl_2Br are shown in Figure 2).

Calibration and linearity

Five levels of concentration were tested in triplicate; these concentrations covered the concentration ranges expected for four halo-organic compounds in water samples (22). Calibration curves were evaluated as:

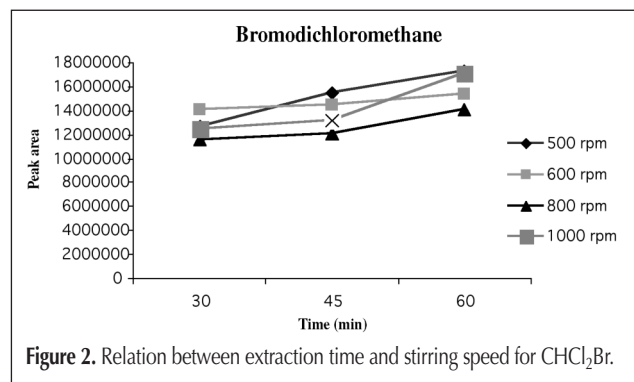
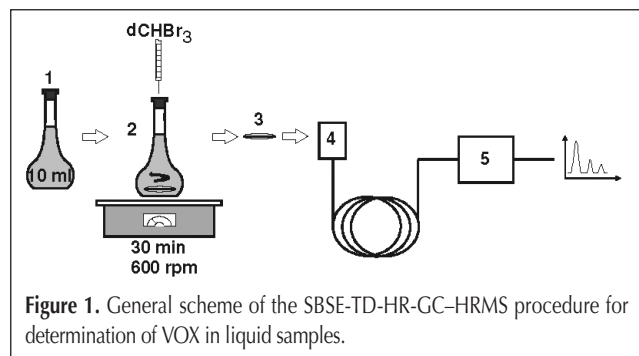
$$Y = f(x)$$

$$P_{\text{VOX}}/P_{\text{CDBr}_3} = f(C_{\text{VOX}}/C_{\text{CDBr}_3}) \quad \text{Eq. 2}$$

where P_{VOX} is the peak height of the compound, P_{CDBr_3} is the peak height of the deuterated CHBr_3 , C_{VOX} is the concentration of the compound added to the sample (0.5, 1, 2, and 5 $\mu\text{g/L}$), and C_{CDBr_3}

Compound	RT* (min)	Selected ions for identification (m/z) [†]
CHCl_2Br	4.4	$\text{M}^+\text{-Br}$ 82.9455 , 84.9426
C_2Cl_4	6.7	M^+ 163.8754 , 165.8725
CHClBr_2	7.1	$\text{M}^+\text{-Br}$ 126.8950 , 128.8927
CHBr_3	9.4	$\text{M}^+\text{-Br}$ 170.8445, 172.8425
Deuterated CHBr_3	9.4	$\text{M}^+\text{-Br}$ 171.8508, 173.8487

* RT = retention time
[†] Ion chosen for quantification in bold.



is the constant concentration of the deuterated CHBr_3 ($2 \mu\text{g/L}$).

The range of concentration, regression line equation, and correlation coefficient appears in Table III. In general, the linearity was very good in the concentration range examined ($0.5\text{--}5 \mu\text{g/L}$) with correlation coefficients greater than 0.979.

LOD and LOQ

The LOD and LOQ were established by considering the mean noise levels on the mass traces chosen for quantification, respectively (Table II). The LOD was set at a signal-to-noise (S/N) ratio of 3 and the LOQ at a S/N of 10, respectively. The LOD and LOQ values obtained in spiked water samples are listed in Table IV. These LOD values are theoretical and show the great potential of this procedure to determine VOC concentrations at very low ppt levels. However, it is necessary to prepare new calibration curves for lower concentration range. Even with a less sensitive benchtop MS, the performance of the method is likely to meet the typical requirements.

Compound	Matrix	Concentration range ($\mu\text{g/L}$)	Regression line equation	Correlation coefficient
CHBrCl_2	Water	0.5–5.0	$y = 0.1721x - 0.0289$	0.9792
C_2Cl_4	Water	0.5–5.0	$y = 0.4521x - 0.0683$	0.9861
CHBr_2Cl	Water	0.5–5.0	$y = 0.3174x - 0.0395$	0.9926
CHBr_3	Water	0.5–5.0	$y = 0.9352x + 0.0366$	0.9940

Compound	LOD ($\mu\text{g/L}$)	LOQ ($\mu\text{g/L}$)
CHClBr_2	0.0017	0.0057
C_2Cl_4	0.00023	0.00076
CHCl_2Br	0.000094	0.00031
CHBr_3	0.000030	0.00099

Compound	Matrix	No. of results	Average $P_{\text{VOX}}/P_{\text{CHBr}_3}$ (for $C_{\text{VOX}}/C_{\text{CHBr}_3} = 1$)	Repeatability (%)
CHCl_2Br	water	5	0.111	8.6
	urine	4	0.104	6.2
C_2Cl_4	water	5	0.321	7.4
	urine	4	0.313	5.1
CHClBr_2	water	5	0.223	9.1
	urine	4	0.286	6.3
CHBr_3	water	5	0.905	10.5
	urine	4	1.05	6.4

Determination of halo-organic compounds in human urine samples

Five urine samples obtained from one volunteer were analyzed using the SBSE-TD-GC–HRMS procedure. One sample was analyzed as a blank sample; the other four were spiked with the same volume of all compounds investigated (to get a concentration of $2 \mu\text{g/L}$) and the same volume of the deuterated CHBr_3 . The repeatability was evaluated on relative $P_{\text{VOX}}/P_{\text{CDBr}_3}$ values using replicates of spiked urine sample, which were analyzed on the same day and by the same analyst. The repeatability for water matrix was evaluated using replicates of five samples. The results show that the obtained $P_{\text{VOX}}/P_{\text{CDBr}_3}$ values did not differ significantly in between the analyses of the urine and water samples (Table V). Values for the repeatability of equal to or lower than 10% allowed the assumption of a good precision for this method.

By applying the isotope dilution technique the concentration of the analytes in the sample (C_{VOX}) are calculated by applying the following equation:

$$C_{\text{VOX}} = (C_{\text{CDBr}_3} \times P_{\text{VOX}}) / (rrf_{\text{VOX}} \times P_{\text{CDBr}_3}) \quad \text{Eq. 3}$$

where P_{VOX} is the peak height of each VOX presented in the sample, C_{CDBr_3} and P_{CDBr_3} are the concentration of the spiked

Compound	Determined C_{VOX}	Theoretical C_{VOX}
CHCl_2Br	1.80	2.0
C_2Cl_4	1.97	2.0
CHClBr_2	1.80	2.0
CHBr_3	1.97	2.0

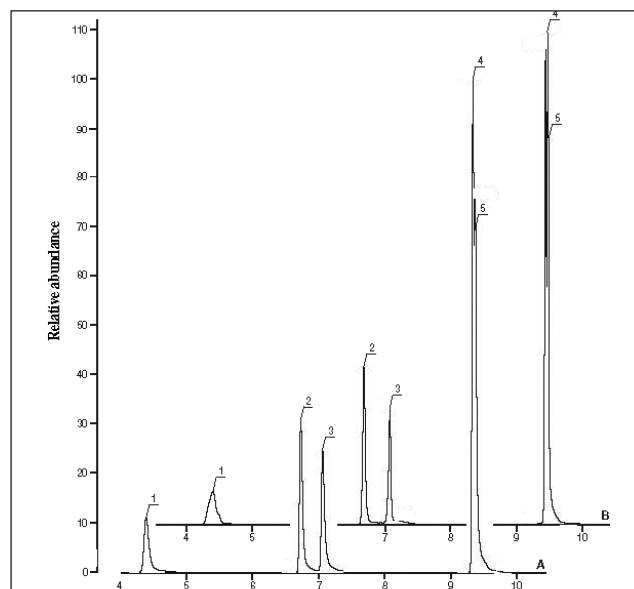


Figure 3. Examples of chromatograms obtained during the analysis of water, (A), and urine samples, (B), with standard solution of volatile halo-organic compounds ($2 \mu\text{g/L}$): CHCl_2Br , 1; C_2Cl_4 , 2; CHClBr_2 , 3; CDBr_3 , 4; CHBr_3 , 5.

labeled standard in the sample and its resulting peak height, and rrf_{vox} is the relative response factor of the analyte in relation to the labeled standard. Rrf_{vox} is calculated from the data of the calibration and linearity experiments by applying the previously rearranged equation. If equation 3 was applied to the spiked urine samples, the results minus the amounts detected in the blank sample should be equal to the spiked concentrations. Concentrations for all compounds presented in blank urine sample (except CDBr_3 , which was added to the urine sample) are below the evaluated concentration range.

The data in Table VI illustrate that the calculated C_{vox} are close to the theoretical spiked concentrations, which confirm that the analytical procedure is performing correctly and the response factors, which were determined in the water matrix, are also valid for the urine matrix. All $P_{\text{vox}}/P_{\text{CDBr}_3}$ values for water and urine standard solutions were statistically compared for all analytes. It shows that little or no matrix effect is observed.

In Figure 3, the chromatogram of a water and a urine sample are shown. Also, no matrix effect on the chromatographic performance could be detected, except a slight peak broadening for CHClBr_2 (little matrix effect).

Conclusion

The procedure of simultaneous determination of four volatile organohalogen compounds (CHCl_2Br , CHClBr_2 , CHBr_3 , and C_2Cl_4) in small samples of water and human urine based on application of SBSE-TD-GC-HRMS has been developed and validated. The key parameters (time, stirring speed) of the extraction step have been optimized to obtain an isolation and concentration method adequate for compounds of high volatility. The results show that the optimum extraction time is 30 min with 600 rpm of stirring speed with a Twister (20 mm in length \times 1.0-mm film thickness) of PDMS (126 μL). The SBSE-TD-GC-HRMS procedure is sensitive and shows a good linearity between 0.5 and 10 $\mu\text{g/L}$ for all compounds tested. This procedure is characterized by very low LOD and LOQ: lower than 0.0017 $\mu\text{g/L}$ and good repeatability for all four volatile compounds. No significant or little matrix effects were observed.

The examined analytical procedure is easy to handle, solvent-free, fast, and was successfully applied for the simultaneous determination of volatile trace compounds in very low volumes of water and human urine samples. Results show the great potential of the SBSE-TD-GC-HRMS procedure with isotope dilution technique used on sampling preparation step can be an excellent alternative to standard HS-GC-ECD-MS procedure.

Acknowledgments

Acknowledge of the LLP-Erasmus programme for financial support.

References

- H. Ukai, S. Inui, S. Takada, J. Dendo, J. Ogawa, K. Isohe, T. Ashida, M. Tamura, K. Tabuki, and M. Ikeda. Types of organic solvents used in small- to medium-scale industries in Japan; a nationwide field survey. *Int. Arch. Occup. Environ. Health* **70**: 385–392 (1997).
- G. Aggazzotti, G. Fantuzzi, G. Predieri, E. Righi, and S. Moscardelli. Indoor exposure to perchloroethylene (PCE) in individuals living with dry-cleaning workers. *Sci. Total Environ.* **156**: 133–137 (1994).
- K. Furuki, H. Ukai, S. Okamoto, S. Takada, T. Kawai, Y. Miyama, K. Mitsuyoshi, and Z.W. Zhang, Higashikawa, and M. Ikeda. Monitoring of occupational exposure to tetrachloroethene by analysis for unmetabolized in blood and urine in comparison with urinalysis for trichloroacetic acid. *Int. Arch. Occup. Environ. Health* **73**: 221–227 (2000).
- N. Paaso, J. Peuravuori, and K. Pihlaja. Extraction efficiency of chloroethenes from contaminated dry cleaner's sludge with three different methods. *Waste Manage.* **20**: 69–74 (2000).
- http://en.wikipedia.org/wiki/International_Agency_for_Research_on_Cancer
- S. Hellweg, E. Demou, M. Scheringer, T.E. McKone, and K. Hungerbuhler. Confronting workplace exposure to chemicals with LCA: Examples of Trichloroethylene in metal degreasing and dry cleaning. *Environ. Sci. Technol.* **39**: 7741–7748 (2005).
- A.J.W. Verplanke, M.H.L. Leumens, and R.F.M. Herber. Occupational exposure to tetrachloroethene and its effects on the kidneys. *JOEM* **41**: 11–16 (1999).
- N. Jakubowska, B. Zygunt, Z. Polkowska, B. Zabiegała, and J. Namiesnik. *J. Chromatogr. A* in press.
- F. David and P. Sandra. Stir bar sorptive extraction for trace analysis. *J. Chromatogr. A* **1152**: 54–69 (2007).
- M. Kawaguchi, R. Ito, K. Saito, and H. Nakazawa. Novel stir bar sorptive methods for environmental and biomedical analysis. *J. Pharm. Biomed.* **40**: 500–508 (2006).
- M. Kawaguchi, K. Inoue, M. Yoshimura, R. Ito, N. Sakui, N. Okanouchi, and H. Nakazawa. Determination of bisphenol A in river water and body fluid samples by stir bar sorptive extraction with in situ derivatization and thermal desorption-gas chromatography-mass spectrometry. *J. Chromatogr. B* **805**: 41–48 (2004).
- E. Baltussen, Sandra P, David F, et al. Stir bar sorptive extraction (SBSE), a novel extraction technique for aqueous sample: Theory and principles. *J. Microcolumn Sep.* **11**: 737–747 (1999).
- A. Stopforth, B.V. Burger, A.M. Crouch, and P. Sandra. Urinalysis of 4-hydroxynonenal, a marker of oxidative stress, using stir bar sorptive extraction-thermal desorption-gas chromatography/mass spectrometry. *J. Chromatogr. B* **834**: 134–140 (2006).
- M. Kawaguchi, R. Ito, Y. Hayatsu, H. Nakata, N. Sakui, N. Okanouchi, K. Saito, H. Yokota, S-I. Izumi, T. Makino, and H. Nakazawa. Stir bar sorptive extraction with in situ de-conjugation and thermal desorption gas chromatography-mass spectrometry for measurement of 4-nonylphenol glucuronide in human urine sample. *J. Pharm. Biomed. Anal.* **40**: 82–87 (2006).
- M. Kawaguchi, Y. Ishii, N. Sakui, N. Okanouchi, R. Ito, K. Saito, and H. Nakazawa. Stir bar sorptive extraction with in situ derivatization and thermal desorption-gas chromatography-mass spectrometry for determination of chlorophenols in water and body fluid samples. *Anal. Chim. Acta* **533**: 57–65 (2005).
- J.P. Lambert, W.M. Mullett, E. Kwong, and D. Lubda. Stir bar sorptive extraction based on restricted access material for the direct extraction of caffeine and metabolites in biological fluids. *J. Chromatogr. A* **1075**: 43–49 (2005).
- B. Tienpont, F. David, T. Benijts, and P. Sandra. Stir bar sorptive extraction-thermal desorption-capillary GC-MS for profiling and target component analysis of pharmaceutical drugs in urine. *J. Pharm. Biomed. Anal.* **32**: 569–579 (2006).
- E.D. Guerrero, R. N. Marín, R. C. Mejías, and C. G. Barroso. Stir bar sorptive extraction of volatile compounds in vinegar: Validation study and comparison with solid phase microextraction. *J. Chromatogr. A* **1167**: 18–26 (2007).
- O. Alvarez-Avilés, L. Cuadra-Rodríguez, F. González-Illán, J. Quiñones-González, and O. Rosarion. Optimization of a novel method for the organic chemical characterization of atmospheric aerosols based on microwave-assisted extraction combined with stir bar sorptive extraction. *Anal. Chim. Acta* **597**: 273–281 (2007).
- E.D. Guerrero, R.C. Mejías, R.N. Marín, and C.G. Barroso. Optimization of stir bar sorptive extraction applied to the determination of pesticides in vinegars. *J. Chromatogr. A* **1165**: 144–150 (2007).
- A. Prieto, O. Zuloaga, A. Usobiaga, N. Etxebarria, and L.A. Fernández. Development of a stir bar sorptive extraction and thermal desorption-gas chromatography-mass spectrometry method for the simultaneous determination of several persistent organic pollutants in water samples. *J. Chromatogr. A* **1174**: 40–49 (2007).
- N. Jakubowska, Z. Polkowska, W. Kujawski, P. Konieczka, and J. Namiesnik. A comparison of three solvent-free techniques coupled with gas chromatography for the determination of trihalomethanes in urine samples. *Anal. Bioanal. Chem.* **388**: 691–698 (2007).

Manuscript received September 13, 2008;
Revision received December 6, 2008.